



Differences in anthropogenic greenhouse gas emissions estimates explained

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29 Abstract

- 30 We examine differences in global and national greenhouse gas (GHG) emissions estimates and highlight the important role of
- 31 varying system boundaries and conceptual approaches in driving these variations. Despite consensus among assessments and
- 32 datasets that GHG emissions continue to increase and are far from aligned with the Paris Agreement goals, estimates can differ
- 33 significantly. Our review finds three main reasons for these differences. First, datasets vary in their coverage of gases, sectors
- 34 and countries; second, there are different approaches to defining 'anthropogenic' emissions and removals in the land use, land-
- use change and forestry (LULUCF) sector; and third, the Paris Agreement doesn't cover all relevant sources of emissions,
- 36 including the cement carbonation sink and ozone depleting substances. As different assessments have different objectives, they
- 37 may deal with these issues differently. We highlight three assessment conventions that report or use emissions data: those
- 38 focused on interpreting national progress, policies and pledges under the Paris Agreement; those consistent with integrated
- 39 assessment modelling (IAM) benchmarks of emissions under different warming scenarios; and those consistent with climate
- 40 forcing assessments. Considering annual average emissions over the period 2014 to 2023, we show global totals of 44.7 GtCO₂e
- 41 yr^{-1} [90% CI ± 4.6], 53.1 GtCO₂e yr^{-1} [90% CI ± 5.2], and 54.9 GtCO₂e yr^{-1} [90% CI ± 5.2] for these three conventions,
- 42 respectively. We suggest that users of GHG emissions data increase transparency in their decision criteria for choosing datasets
- 43 and setting the scope of an assessment. The data used in this study to make figures 8-13 is available at:
- 44 <u>https://doi.org/10.5281/zenodo.15126539</u> (Lamb, 2025b).





45 **1** Introduction

46 A key indicator to assess human influence on the climate is total anthropogenic greenhouse gas (GHG) emissions. At a global

47 level, tracking developments in this metric is necessary to evaluate progress towards the climate objectives of the Paris

48 Agreement - including the human contribution to warming so far, the timing of peak emissions, and how fast reductions need to

49 proceed in the coming decades. At the national level, tracking GHG emissions trends is instrumental to evaluating the climate

- 50 policy progress of countries.
- 51

52 Despite its centrality to climate and climate policy assessments, different communities report different levels and trends in total

53 anthropogenic GHG emissions - even though all assessments show GHG emissions have conclusively increased over the past

54 decades, and are off track from a pathway consistent with the goals of the Paris Agreement (Forster et al., 2024; IPCC, 2022; 55

UNEP, 2024; UNFCCC, 2022c). Nevertheless, as we will show below, global estimates of annual GHG emissions can vary by a

56 margin of several GtCO₂e. This is a phenomenon that has received increasing attention in the literature and in the 57

Intergovernmental Panel on Climate Change (IPCC) (Gidden et al., 2023; Grassi et al., 2023; IPCC, 2024). These differences are

58 often related to different input datasets, different definitions and scope, as well as decisions regarding what is included in the

59 estimates (Andrew, 2020), in addition to the underlying uncertainties.

60

It is important to distinguish differences in emissions reporting that result from different methodologies and data sources, which 61

62 can represent uncertainty and data quality in our understanding, versus those that result from alternative conceptual approaches

63 and system boundaries. An example of the former would be the use of different emissions factors (EFs) across datasets or data

64 versions, which provide an estimate of the emissions associated with a given activity. An example of the latter is the fact that

65 assessments may choose to exclude certain emissions categories, for instance from biomass fires, while others include them. In

66 some cases, it may not be made explicit that an emission source is excluded in an assessment. In this article we are concerned

67 with the latter kind of decisions and the fact that specific system boundary choices greatly matter for tracking GHG emissions. 68

69 There is no single agreed approach to setting the system boundaries of an emissions assessment. Even if the same input data and 70 emission factors are used, different communities have developed their own conventions on which categories of emissions are

71 included in an assessment. Two of the main communities include users of national greenhouse gas inventories (used for country

72 reporting), and the scientific communities performing climate and integrated assessment modeling (IAM). Further, national,

73 regional and global GHG emissions are widely depicted in a variety of reports and the decision criteria for which components of

74 emissions are included or excluded are often poorly transparent (Boehm et al., 2023; European Commission, 2024; Forster et al.,

75 2024; UNEP, 2024; UNEP and CCAC, 2021; USGCRP, 2023).

76

77 It is critical to explain the decision criteria behind system boundary choices in emissions reporting, and to understand the

78 consequences of these differences. Besides the fact that different published estimates lead to general confusion among non-

79 domain experts, this issue can compromise important science-policy processes. For example, differences in land use, land-use

80 change and forestry (LULUCF) emissions between national inventories and IAMs are highly consequential for calculating

81 benchmarks to meet the 1.5°C and 2°C goals, including when countries should reach net-zero or if calculated net-zero targets

82 would be enough to stabilise global temperatures (Allen et al., 2025; Gidden et al., 2023; Grassi et al., 2021). Further, there is a

83 risk that observers start to lose trust in emissions estimates, including the official inventories published by countries, simply

84 because they are perceived to misrepresent or exclude certain sources (Mooney et al., 2021; Yona, 2025).

85

86 In this perspective we ask three questions. First, what are the main system boundary issues causing GHG estimates to differ?

87 Second, what conventions are taken in different assessment communities with respect to these system boundaries? And third,

88 what is the possible spread in global or national GHG estimates according to these conventions? In answering these questions,

89 we aim to explain and promote transparency in key decision criteria that lie behind GHG emissions assessments. In terms of

90 scope, our discussion covers the main well-mixed GHGs that are covered by the Kyoto Protocol, the Paris Agreement and the

91 Montreal Protocol, including carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O) and Fluorinated gases (F-gases)

92 including Ozone Depleting Substances (ODS). We do not extend our analysis to other climate relevant emissions (e.g. SOx,

93 NOx, CO, etc.), while recognizing that these too have relevant impacts on atmospheric chemistry and the climate. We also do not

94 consider the role of global warming potential metrics, even though different choices here can obviously lead to varying

95 estimates.





96 2 Three reasons why greenhouse gas estimates can differ

97 Emissions inventories form the basis for most national and global reporting of anthropogenic GHG emissions. These 'bottom-up'

- 98 accounts are constructed by tracking human activities in different domains (e.g., fuel use, cement production, land use
- transitions, livestock numbers) and estimating the expected GHG emissions or removals under different (e.g. technology or
- climate) conditions. Combustion emissions are usually estimated by multiplying fuel use by the corresponding emission factor.
- Some of these accounts also depend on modelling, particularly for agricultural and land-based activities. A number of different
- 102 datasets are now available and are in widespread use across the climate research community to estimate global and national GHG
- 103 emissions. However, despite being well documented, there are several key reasons why studies using them can arrive at quite
- 104 different estimates of global or national emissions.

105 2.1 Datasets vary in their coverage of gases, sectors and countries

Bottom-up datasets generally aim to cover the set of emissions sources outlined by the United Nations Framework Convention on Climate Change (UNFCCC, 2018b). In terms of gases, this includes CO₂, CH₄, N₂O and a subset of F-gases covering

108 hydrofluorocarbons (HFCs), Perfluorocarbons (PFCs), Nitrogen Trifluoride (NF₃) and Sulfur Hexafluoride (SF₆). These are often

109 referred to in the literature as the "Kyoto gases". In terms of sectors, national reporting to the UNFCCC includes five main

110 categories (and many more subcategories) which sum to the total: 1. Energy, 2. Industrial process and product use (IPPU), 3.

111 Agriculture, 4. LULUCF, and 5. Waste.

112

113 In principle, the national GHG inventories that countries submit to the UNFCCC cover all of these sources, and would enable 114 complete assessments of global emissions if all countries submitted regularly over time. Countries are also guided to submit an

annual time series starting in 1990, reporting emissions up to two years prior, which would cover many use cases (UNFCCC,

- 116 2018b). However, in practice the UNFCCC recognises that countries have different capabilities and statistical infrastructures,
- and has made a series of allowances for non-Annex I countries: they have the flexibility to report three gases (CO₂, CH₄, N₂O)

instead of all seven (i.e. they may exclude F-gases); they may report annual time series from 2020 onwards instead of from 1990;

and they may do so up to three years prior instead of two (UNFCCC, 2018b §48, §57 & §58). The Biennial Transparency

120 Reports have improved the frequency of inventory reporting, but still make allowances for Least Developed Countries and Small

- 121 Island Developing States to submit at their discretion.
- 122

123 As a result, national GHG inventories are typically only complete and timely for Annex I countries from 1990 onwards, covering 124 about one-third of total world emissions. Depending on the use case, this can be too restrictive, meaning that third-party datasets 125 produced by researchers and International Institutes alike are frequently used instead to report global or national totals, including 126 trends before 1990. These third-party datasets usually explicitly follow the inventory convention for sectors and coverage of 127 gases, but make use of national statistics for activity data and independently assessed emissions factors, often based on general 128 default values (also known as "Tier 1" estimates). Key global datasets that cover multiple sectors and gases with a global scope 129 include: the Emissions Database for Global Atmospheric Research (EDGAR) (Crippa et al., 2024; Janssens-Maenhout et al., 130 2019); the Community Emissions Data System (CEDS) (Hoesly et al., 2025); and the Potsdam Realtime Integrated Model for 131 probabilistic Assessment of emissions Paths (PRIMAP-Hist) (Gütschow et al., 2025). Sectorally-focused datasets include the 132 Global Carbon Project's (GCP's) Global Carbon Budget (GCB) (Friedlingstein et al., 2025), Global Methane Budget (GMB) 133 (Saunois et al., 2024) and Global Nitrous Oxide Budget (GNB) (Tian et al., 2024); the Energy Institute's Statistical Review of 134 World Energy (EI - formerly published by BP) (Energy Institute, 2025); the International Energy Agency (IEA) GHG Emissions 135 from Energy dataset (IEA, 2024), and the Food and Agriculture Organisation of the UN (FAOSTAT) Greenhouse Gas Emissions 136 dataset (FAOSTAT, 2025). Several further useful datasets include the gap-filled and harmonised dataset of LULUCF data based

137 on National GHG inventories (Grassi et al. NGHGI) (Grassi et al., 2022a, b); the Greenhouse Gas and Air Pollution Interactions

and Synergies (GAINS) dataset of methane emissions (Höglund-Isaksson et al., 2020); the Global Fire Emissions Database

139 (GFED), the Global Fire Assimilation System (GFAS), and the Global Wildfire Information System (GWIS) for fire emissions

140 (Giglio et al., 2013; GWIS, 2025; Kaiser et al., 2012; van der Werf et al., 2017); Andrew (2025) for cement emissions; and

- 141 inversion datasets for ODS and F-gas emissions (Forster et al., 2024; Velders et al., 2015; WMO, 2022).
- 142

143 The appropriate use of these datasets is complicated by several issues. The first is that, as it stands, no single third-party dataset

has complete and up-to-date coverage of all UNFCCC relevant gases, sectors and countries (Table 1). Only three datasets cover

145 all GHGs in the convention (EDGAR, PRIMAP-Hist and FAOSTAT) and while many cover agriculture, most exclude LULUCF

146 emissions. Only two datasets cover global emissions of non-CO₂ LULUCF emissions (FAOSTAT, PRIMAP-Hist). To obtain a





147 complete global or national total across all gases it is therefore often necessary to combine multiple datasets, for example by 148 using EDGAR in combination with CO₂ LULUCF from GCB.

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150 Second, due to different formats, overlaps between datasets, and varying methodological approaches, it is generally advised to

151 take care when combining them. Each dataset is not necessarily like-for-like; EDGAR for example often applies global average emission factors (Tier 1) while most UNFCCC Annex I countries apply national emission factors and/or models (Tier 2 or 3), 152

153 which can lead to differences in emission levels and trends, particularly at the national level and for non-CO₂ GHGs. Further,

154 individual datasets have different approaches to dealing with certain sectors. For instance, in the case of emissions from bunker

155 fuels (international aviation and shipping): the national GHG inventories report these as a memo item for each country (i.e.,

156 excluded from national totals); EDGAR and GCB report these as a single stand-alone "country"; EI includes these in national

157 totals; CEDS reports these as a sector in the global total; and PRIMAP-Hist excludes them entirely. Differences in accounting for emissions and removals in the LULUCF sector are even more consequential, as discussed in the next section.

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159

160 Third, there can be significant dependencies between datasets (Andrew, 2020), for example many datasets rely ultimately on

161 activity data reported by the IEA, the UN Statistics Division (UNSD), EI and FAOSTAT. The PRIMAP-Hist dataset is a prime

162 example of this, being an amalgamation of several underlying data products, with two individual time series: the "CR scenario",

163 which prioritises national GHG inventory data and gap fills these with third-party data (EI, Andrew, FAO, EDGAR); and the

164 "TP scenario", which prioritises the latter third-party data. Conversely, the FAOSTAT GHG emissions dataset has begun to

- 165 incorporate PRIMAP-Hist data for energy, IPPU and waste emissions. Changes in underlying datasets can therefore cascade across many of the datasets we discuss here.
- 166 167

	Reference	Gases	Sector coverage							
Dataset			1.A. Energy (fuel combustion)	1.B. Energy (fugitive emissions)	1.D.1. Energy (Intl. Bunkers) [†]	2. Industrial process and product use	3. Agriculture	4. Land use, land use change and forestry	5. Waste	Reporting frequency and delay
UNFCCC Inventories*	UNFCCC (2025)	CO ₂ , CH ₄ , N ₂ O, F-gases	yes	yes	yes	yes	yes	yes‡	yes	Annual, 2 year delay (Annex 1)
EDGAR	Crippa et al. (2024)	CO ₂ , CH ₄ , N ₂ O, F-gases	yes	yes	yes	yes	yes	yes (for regions) [‡]	yes	Annual, 1 year delay
IEA	IEA (2024)	CO ₂ , CH ₄ , N ₂ O	yes	yes	yes	some fluxes	no	no	no	Annual, 1 year delay
CEDS	Hoesly et al. (2025)	CO ₂ , CH ₄ , N ₂ O	yes	yes	yes	yes	yes	no	yes	Annual, 1 year delay
GCB	Friedlingstein et al. (2025)	CO ₂	yes	yes	yes	most fluxes [¶]	most fluxes¶	yes‡	yes	Annual, 1 year delay
EI (BP)	Energy Institute (2025)	CO ₂ , CH ₄	yes	yes	yes [§]	some fluxes#	no	no	some fluxes	Annual, 1 year delay
PRIMAP- Hist	Gütschow et al. (2025)	CO ₂ , CH ₄ , N ₂ O, F-gases	yes	yes	no	yes	yes	yes‡	yes	Annual, 1 year delay
FAOSTAT	FAOSTAT (2025)	CO ₂ , CH ₄ , N ₂ O, F-gases	yes	yes	yes	yes	yes	yes‡	yes	Annual, 2 year delay

168

169 Table 1: Bottom-up emissions datasets of anthropogenic GHG emissions and their characteristics. Several subcategories of the energy

170 sector are shown to highlight their exclusion in some datasets. Datasets are named and referenced in section 2.1. Notes: * Only Annex I

171 countries reliably submit complete inventories each year. [†] Bunkers are included as a memo item in UNFCCC inventories (excluded from

172 national totals), and typically as a separate "country" in other datasets. ‡ Definitions of LULUCF differ, as discussed in section 2.2. ¹ For some

173 countries, excludes lime, glass and other decomposition in section 2, and liming in sector 3. § Included in national totals and not reported 174

separately. # Includes cement only. Adapted from (Andrew, 2020). Note that PRIMAP-Hist includes two datasets (Hist-CR and Hist-TP), 175

which prioritise data from national inventories and third-party sources, respectively. Red colours indicate incomplete coverage but do not

176 indicate how important this is for the total assessment of emissions (e.g. in GtCO2e).





177

- 178 A further complication is that dataset methodologies can carry implicit system boundary decisions. For instance, the IPCC
- 179 Guidelines on National Greenhouse Gas Inventories (IPCC, 2006, 2019), used by countries to calculate, format and submit data
- to the UNFCCC, recommend at Tier 1 the use of fuel sales data to calculate road transport emissions. Since fuel use is
- transboundary in nature, this means that large discrepancies can be observed between the Tier 1 inventory approach and higher
- 182 Tiers datasets that apply a more refined territorial principle (i.e. using modelling studies to estimate fuel consumption within a 183 country) (e.g. BMK, 2023). Likewise, the IPCC Tier 1 methodology for Harvested wood products (HWPs) follows the so-called
- production approach and explicitly assumes that end-of-life emissions from traded HWPs occur within the country from which
- they were exported, rather than in the importing country (IPCC, 2019). Assuming countries all follow the same principle, at the
- global level these differences balance out, and the IPCC Guidelines methodological choices were likely made in the interests of
- 187 both pragmatism and to reduce the chances of omissions or double-counting at aggregated levels.
- 188
- 189 Together, these issues mean that dataset choices matter, and that assessments often have to combine different datasets to gain
- 190 totals that are comparable to the scope of national GHG inventories. Further, this requires caution due to potential overlaps and
- 191 conceptual differences between datasets.

192 **2.2** There are different approaches to defining 'anthropogenic' emissions

193 A second issue affecting comparability in emissions assessments is that different communities and datasets have different

- 194 approaches to estimating or even defining 'anthropogenic' emissions and removals (together: fluxes). Specifically, this issue
- arises in connection with greenhouse gas fluxes in terrestrial ecosystems (e.g. forests and wetlands), which can be influenced by
- three main types of effects: (1) direct anthropogenic effects, such as changes in land use (e.g. deforestation or crop abandonment)
- and various types of management practices; (2) indirect anthropogenic effects, which include environmental changes caused by
- humans, like alterations in temperature, precipitation, CO₂ levels, and nitrogen deposition which can impact growth rates,
- 199 mortality, decomposition, and natural disturbance patterns; and (3) natural effects, including climate variability and inherent
- 200 natural disturbances such as fires and pests (Grassi et al., 2021; IPCC, 2019). The difficulty arises with the second category of
- 201 'indirect anthropogenic' effects, such as when increased atmospheric CO₂ concentrations influence forest growth, or when fires
- burn with an intensity and frequency that would be unlikely without climate change. Since these emissions would not occur without human activity, they cannot be treated as purely natural sources. However, at the same time they do not carry the same
- degree of human intent or direct influence as, for example, the combustion of fossil fuels or the logging of forests. Hence they
- are often given the terms "indirect anthropogenic effects" or "human-induced environmental changes" in the literature (e.g.
- 206 Houghton et al., 2012).
- 207

208 The separation of direct anthropogenic, indirect anthropogenic and natural sources of emissions is conceptually challenging and

- 209 can be difficult to communicate to users of emissions data. Here we cover three of the main areas of emissions accounting where
- 210 indirect anthropogenic effects arise, all of which occur in the LULUCF sector of national inventories, but also to some degree in
- 211 the agriculture sector.

212 2.2.1 Forest land CO₂

- 213 CO₂ emissions and removals on forest land in the LULUCF sector occur are generally complex, difficult to track and involve
- 214 significant uncertainties, but are nonetheless highly consequential for global estimates of GHG emissions. There are two main
- approaches to account for forest land fluxes: the approach developed by the IPCC guidelines and implemented in the national
- 216 GHG inventories and the FAOSTAT emissions dataset, and the global bookkeeping model approach, such as the one
- 217 implemented by the Global Carbon Project. Both track changes between different types of land use (e.g. forest land, cropland,
- 218 grasslands, settlements, other land) and how they influence various carbon stocks (e.g. living biomass, soil organic matter, etc.).
- However, they differ conceptually in one important respect: how they estimate the anthropogenic component of emissions and
- 220 removals (Grassi et al., 2018, 2021, 2023; IPCC, 2024; Schwingshackl et al., 2022).
- 221
- 222 The national GHG inventory approach is primarily survey-based and pragmatically counts all fluxes on "managed land" as
- 223 anthropogenic, including both direct and indirect anthropogenic effects. Simply put, countries estimate in line with national
- definitions which areas of land are 'managed' in their inventories; track this area consistently over space and time; and
- 225 compute the resulting fluxes as anthropogenic. All other areas and fluxes are treated as unmanaged and hence natural. This





226 convention came about not because of political convenience, but because direct observations cannot easily separate out direct

227 anthropogenic and indirect anthropogenic effects (Pongratz et al., 2021). A consequence of the inventory approach is that the

quantified fluxes depend critically on the definition of "managed land". Conventionally, "managed land" is defined in a broad sense to include land that "perform[s] production, ecological, or social functions" (IPCC, 2006, 2024). In addition to cropland,

and managed forests, this may include large areas of national parks, indigenous lands or areas subject to fire-protection activities,

and managed forests, this may include large areas of national parks, indigenous lands of among others.

232

Even within the national GHG inventory approach, important differences may arise due to use of different IPCC tier methods, despite similarities in the land use approach. These include the Tier 1 'gain-loss' method (estimating fluxes due to deforestation,

harvest, regrowth, etc.) and the Tier 3 'carbon stock change' approach (deriving fluxes from changes in biomass stocks over

time). While many country inventories, including those used as input into the Grassi NGHGI dataset, apply gain-loss, the

FAOSTAT forest data in LULUCF are estimated using the carbon stock change approach, using country data from the FAO

Forest Resources Assessment (Tubiello et al., 2021). The application of one of the two methods also brings differences in input

datasets and their quality, and is responsible for most of the differences produced as a result (Fig. 1). Additional important

differences between Grassi NGHGI and FAOSTAT stem from more complete coverage of sources in the former, including soils

stocks, whereas the latter is limited to estimating carbon stock changes in above-ground biomass (Grassi et al., 2022a). In

particular, the forest sink is underestimated in FAOSTAT compared to Grassi NGHGI, because the underlying carbon stock data

from many developing countries are incomplete.

244

245 By contrast, bookkeeping models quantify fluxes that are the result of direct human intervention (e.g. deforestation, harvest,

regrowth) and exclude, by simulation, those that are natural responses to human-induced environmental changes (i.e. indirect

247 effects) - the most important of which is the increase in vegetation growth due to rising atmospheric CO₂ concentrations. The

248 bookkeeping approach is independent of definitional choices related to the managed land area, as it distinguishes natural from

anthropogenic fluxes not by area, but by driver (whether or not there is land-use activity). This means that implicitly all land is

considered, independent of a definition of being managed or not, but fluxes only occur when land management or land-use

- change as defined by the models take place.
- 252



Differences in net land use change (LULUCF) estimates

Figure 1: Differences in net land use change (LULUCF) estimates. Data: bookkeeping models BLUE, OSCAR, H&N and their average
 from Friedlingstein et al. (2025); FAOSTAT from FAOSTAT (2025) and Tubiello et al. (2021); Inventories from Grassi NGHGI (Grassi et al.,
 2022a). Composite data based on inventories and FAO are also available in PRIMAP-Hist, but not shown here.

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253

The result of these conceptual differences is a gap of about 7.5 $GtCO_2 yr^{-1}$ between inventory and bookkeeping estimates of

LULUCF CO₂ emissions (10 year average up to 2022). This has diverse consequences for global benchmarks of mitigation

action (Gidden et al., 2023; Grassi et al., 2021), and could also have important equity implications. Thus the simple choice of

261 national GHG inventory versus GCB data for the LULUCF sector - keeping everything else constant - can significantly affect





- 262 global GHG estimates and thus greatly affect mitigation scenario analysis. To this regard, the IPCC (2025a) has indicated in the
- 263 outlines of its AR7 reports that estimates and scenarios for human-induced, land-based CO2 fluxes will need to consider
- 264 alignment with national inventory definitions.

265 2.2.2 Natural disturbances

266 Fires and other disturbances occur on land, including the managed lands covered by national GHG inventories, and can generate

267 significant emissions of CO₂, CH₄ and N₂O. To illustrate, an estimated 8.8 GtCO₂ was released in March 2023 – February 2024

268 fire season including extreme wildfires in Canada that were around 3 times more likely due to anthropogenic climate change 269 (Jones et al., 2024b). In a stable fire regime, the vegetation on burnt areas generally recovers in subsequent years, drawing down

270

CO2 from the atmosphere during the recovery phase. In principle, this suggests that fire emissions could have a net zero impact 271

on atmospheric CO₂ emissions over multiple decades under a natural fire regime. However, observed increases in the extent and 272 severity of fires under climate change point to shifts in fire regimes that ultimately lead to more disturbed landscapes that store

273 less carbon (Cunningham et al., 2024; Jones et al., 2024a).

274

275 The key problem with fires is that although they can occur naturally, they are now more likely than in the pre-industrial period

276 due to anthropogenic climate change. This leads to major definitional obstacles to separating "anthropogenic" from "natural" fire

277 emissions. The distinction is further complicated by the mixture of anthropogenic and natural (lightning) ignitions that occur.

278 Anthropogenic ignitions are themselves complicated because some fires are deliberately set to clear land for agriculture or for

279 land management purposes or for arson, whereas others are unintentional (e.g. power infrastructure failure or dropped cigarettes).

280 Today's anthropogenic ignitions must also be viewed within the context of historical rates of ignition by people, which is

281 challenging due to poor constraints on pre-industrial fire use which lead to unreliable counterfactuals. These many complications 282 lead to different interpretations of how fires should be accounted for in global GHG budgets. Despite this, there is a large

283 literature and community studying fires, and numerous satellite-driven observational datasets that are used across different

284 approaches (Giglio et al., 2013; GWIS, 2025; Kaiser et al., 2012; van der Werf et al., 2017).

285

286 In the national GHG inventories, including FAOSTAT (Prosperi et al., 2020), CO₂, CH₄ and N₂O emissions from fires and

287 controlled burning (e.g. of crop residues, prescribed burning, and savannahs) are reported in the agriculture and LULUCF sectors

288 (with the CO₂ emissions from some components disregarded on the assumption of equivalence between emissions and

289 removals). In the LULUCF sector, countries either explicitly report burnt areas and their associated emissions, or they implicitly

290 report these events in their net account of forest biomass stock changes. As discussed in the prior section, the inventories follow

291 an area-based approach and account for all types of fires on managed land - whether they were ignited by anthropogenic or

292 natural means, and regardless of how anthropogenic climate change has influenced their odds. In terms of reporting however, 293

there is an important difference: in accordance with decision 18/CMA.1 §55 (UNFCCC, 2018b), countries may choose to report 294 'wildfire' events as a "natural disturbance" memo item, and exclude the associated emissions and subsequent removals from

295 their national totals and related climate targets. So far, Australia and Canada have made use of this convention in their

296 inventories, significantly altering the sum total of reported annual anthropogenic emissions estimates and their trends (Fig. 2). In

297 principle, the underlying assumption in this rule is that these natural disturbance exclusions would be carbon neutral with respect

298 to subsequent regrowth in post-fire years.







Figure 2: Alternative approaches to accounting for wildfires in national GHG inventories. Australia and Canada report wildfires as the memo "natural disturbances on managed land" and exclude these emissions and subsequent removals from their totals. Note that Canada started to count natural disturbances before 1990 and therefore has excluded removals in the early 1990s that occurred on previously burnt areas. Other countries have so far not used the natural disturbances memo and instead report and account for wildfires on managed land in the LULLUCF sector, even in years with major events - such as Portugal in 2017. Data: National GHG inventories compiled by Lamb (2025a).

305

306 By contrast, the GCB takes a strict interpretation of anthropogenic, with CO2 emissions from fires included implicitly in the 307 estimate of land-use change fluxes as part of the fluxes representing fast release of carbon to the atmosphere (as opposed to 308 slower decomposition of material on site or as products). As the GCB defines land-use change fluxes by driver (land-use 309 activity), these fires are often related to deforestation and shifting cultivation activities that free up land for anthropogenic use. 310 Emissions from wildfires related to anthropogenic global warming or vegetation productivity changes are not considered as part 311 of the land-use change fluxes, but rather as an emission term in the land sink. However, a change in climate may increase the 312 odds that agricultural management or forest clearing fires escape and have a larger than 'intended' effect. This can be observed 313 in, for example, high land-use emission estimates associated with peat drainage and fires in dry El Niño years. These synergistic 314 terms of direct and indirect drivers are included in the GCB land-use change fluxes as part of peat drainage and peat fire 315 emissions. Problematically, the poor representation of the spatial distribution and trends of global fire by dynamic global 316 vegetation models (Jones et al., 2022; Kloster and Lasslop, 2017), as well as major fire emissions anomalies such as that by 317 Canada's wildfires of 2023 or Australia's Black Summer bushfires of 2019/20, leads to missing fluxes of CO2 in estimates of the 318 global land sink and likely contributes to imbalance in the global budget (Friedlingstein et al., 2025; Sitch et al., 2024). 319







Differences in fire emissions accounting

(while noting that some ignitions are natural).

320
 321 Figure 3: Differences in fire emissions accounting. Note: box areas are not representative of total fluxes in each component.

disturbances.

322

A third approach to accounting for fire emissions is represented in the Global Methane and Nitrous Oxide Budgets (Saunois et

al., 2024; Tian et al., 2024), as well as in the FAOSTAT approach (Prosperi et al., 2020) which as in the other cases typically

system boundary and classify all CH, and

N,O emissions from fires as anthropogenic

'natural'

325 draw from satellite-driven observational datasets such as GFED or GFAS but do not didistinguish between anthropogenic and

- natural fires, nor between managed and non-managed land areas (although the GFED database categorises fires as agricultural,
- 327 deforestation and other types). In the Methane and Nitrous Oxide budgets these are known as "biomass fires" and to date have
- 328 simply been accounted as fully anthropogenic in the totals. Total CH₄ and N₂O emissions from fires are significant at at least 0.5
- 329 GtCO₂eyr⁻¹ but with a highly variable trend (Fig. 4).
- 330







331

Figure 4: Global methane and nitrous oxide emissions from fires. Data: GFEDv4.1 (Giglio et al., 2013). CO₂e emissions are calculated
 using GWP100 from AR6 WGI Chap. 7 (Forster et al., 2021).

334 2.2.3 Wetlands and freshwater body emissions

A range of different landscapes and land use types produce large quantities of CH_4 emissions via the anaerobic decomposition of organic matter. These include wetlands (e.g. peatlands, bogs, marshes) as well as freshwater bodies (reservoirs, canals, ponds, etc.). For both of these together, the GMB reports decadal (2010 to 2019) average emissions of 248 [159 to 369 min-max]

338 MtCH₄ yr⁻¹ or 6.9 [4.4 to 10.3] GtCO₂e yr⁻¹ (Saunois et al., 2024). In addition, N₂O emissions are produced by peatland drainage. 339 However, as in other areas of land use, there are major definitional obstacles to determining the anthropogenic component of

these emissions, for example due to nutrient runoff into natural systems, as well as the influence of climate change on them.

341

342 In the national GHG inventories, methane emissions from wetlands and freshwater bodies both fall under the LULUCF

343 "wetlands" sector and are restricted to managed areas of peatlands and flooded lands resulting from artificially constructed water

bodies such as reservoirs, canals, ditches and ponds. The IPCC guidelines consider that emission changes on managed peatland

are insignificant due to drainage, effectively excluding these emissions. For flooded lands, the IPCC guidance now covers

- 346 methane emissions resulting from constructed water bodies and provides default emission factors based on latitudinal region to
- be applied to the created flooded surface. As in the case of forest land fluxes and natural disturbances, emissions from
 unmanaged wetlands are not estimated nor included. FAOSTAT covers only N₂O emissions from peatland drainage (Conchedda
- 349 and Tubiello, 2020).
- 350

The Global Methane Budget, and the wider scientific community estimate wetland and freshwater body emissions separately

based on biogeochemical models driven by the so-called wetland extent. Major uncertainties arise from difficulties in

determining the extent of these areas, for example because they are under vegetated cover, or because they are in close proximity

- to other ecosystem types. Individual studies have estimated global emissions from reservoirs (e.g. Harrison et al., 2021; Johnson
- et al., 2021), rivers and streams (e.g., Rocher-Ros et al., 2023) and lakes and ponds (e.g. Johnson et al., 2022; Zhuang et al.,
- 356 2023), which are classified as inland freshwater ecosystems in the Global Methane Budget. As these studies do not necessarily 357 distinguish the anthropogenic component, an attempt was made to do so in the latest Global Methane Budget (Saunois et al.,
- 2024) and suggested that about half (56 of 112 Tg CH₄ yr⁻¹ or 1.6 GtCO₂e yr⁻¹) of the freshwater emissions indirect
- anthropogenic. This considered artificially constructed water bodies such as reservoirs and farmer ponds, as well as indirect
- 360 anthropogenic disturbances such as eutrophication, erosion and runoff of agricultural landscapes, as well as warming.
- 361

362 Wetland emissions are considered as natural in the GMB, even though some could be considered anthropogenic systems (e.g.

restoration activities) and most are subject to indirect effects via climate change. About 30 Tg of 159 Tg CH₄ yr⁻¹ or 0.8 GtCO₂e yr⁻¹ of wetland emissions are considered as anthropogenic disturbances (Saunois et al., 2024). Since few studies have estimated

- 364 yr⁻¹ of wetland emissions are considered as anthropogenic disturbances (Saunois et al., 2024). Since few studies have estimated 365 anthropogenic disturbances of wetland and inland freshwater emissions, such values should be taken with caution. As in the case
- of fires, these emissions are climate sensitive (through temperature and moisture) and warming has already led to increased
- 367 methane emissions from wetlands as calculated by biogeochemical models (Zhang et al., 2025).







Figure 5: Methane emissions in countries with high shares of wetland emissions. Data: National GHG inventories compiled by Lamb
 (2025a).

372

373 Even though global indirect CH₄ fluxes from wetlands and freshwater bodies are assessed to be large at approximately 2.4

374 GtCO₂e yr⁻¹, national GHG inventories - which are restricted to "artificial" water bodies on managed lands - account these

emissions as very small. Figure 5 shows the CH₄ inventories of six countries with the highest shares of the wetlands sector in

their total CH₄ estimates. With the exception of Iceland (>50% share), these emissions are trivial compared to the livestock,

377 waste or fuel production sectors. Most countries stand at well below 1% of their total CH₄ emissions from wetlands.

378 2.3 The Paris Agreement does not cover all relevant sources of emissions

The third reason why GHG estimates can differ is that current UNFCCC guidance does not cover all climatically relevant sources of emissions and removals. This stems from the existence of other global environmental agreements and the fact that

inventory reporting guidance is not as agile in updating its scope compared to the wider literature. And since inventories exclude

certain sources and gases, this has a knock-on effect on third-party datasets that harmonise with the UNFCCC approach. Two

383 major current omissions are ozone depleting F-gases and the cement carbonation sink.

384 2.3.1 Ozone Depleting Substances (ODS): F-gases

385 Fluorinated gases are human made substances that are widely used in industrial processes and consumer products, for example as

refrigerants, aerosols, and insulation materials. F-gases have high global warming potentials, which were assessed in the IPCC 6th Assessment Report (Forster et al., 2021).

388

389 National inventory reporting and some third-party datasets include estimates of HFCs, PFCs, SF₆ and NF₃. We call these the

- 390 "UNFCCC F-gases". (As mentioned before, the UNFCCC F-gases plus CO₂, CH₄ and N₂O are often referred to as the "Kyoto
- 391 gases"). However, national GHG inventories and the Paris Agreement excludes two further categories of F-gases, namely
- 392 chlorofluorocarbons (CFCs) and hydrochlorofluorocarbons (HCFCs). We call these the Ozone Depleting Substances or "ODS F-





- 393 gases". The ODS F-gases also have high global warming potentials, but were already regulated under the 1987 Montreal
- 394 Protocol and the subsequent Kigali amendment because of their impacts on the ozone layer.
- 395



Figure 6: Total fluorinated (F-) gas emissions. Data: Inversions from Forster et al. (2024). CO₂e emissions are calculated using GWP100
 from AR6 WGI Chap. 7 (Forster et al., 2021).

399

400 The Montreal Protocol was successful in reducing ODS F-gas emissions (Fig 6) and consequently expected levels of global

401 warming (Velders et al., 2007; Young et al., 2021). However, reductions have leveled off in the past decade and there is known

to be a large quantity of these gases in storage and end-use devices that will continue to emit over the coming decades. As it

403 stands, UNFCCC F-gases accounted for approximately 1.8 GtCO₂e yr⁻¹ (90% CI \pm 0.54) in 2022, while ODS F-gases contributed

404 1.5 GtCO₂e yr⁻¹ (90% CI \pm 0.45) (Forster et al., 2024). Thus, while emissions of F-gases can be well estimated using top-down

405 methods - since there are no natural sources and they only break down chemically - a portion of them are not always accounted

406 for in total and national emissions estimates, simply because of their exclusion in UNFCCC reporting.

407 2.3.2 Cement carbonation

408 Atmospheric CO_2 is gradually absorbed into cement materials that are exposed to air, a process known as cement carbonation. 409 This is a slow process over decades, but a globally significant one, because of the enormous quantity of cement that is produced 410 and used in the built environment.

411

The GCB tracks the global cement carbonation sink (Friedlingstein et al., 2025), which itself is based on a bottom-up assessment of cement production and use statistics (Huang et al., 2023). Current estimates indicate a global sink of $0.8 \text{ GtCO}_2 \text{ yr}^{-1}$ that has steadily and rapidly increased alongside cement production. This is currently sufficient to compensate for about one third of

415 cement process emissions (Huang et al., 2023). However, uncertainty is currently large, particularly due to lack of data on the

- 416 share of cement that is used for concrete versus mortar, which are products with very different rates of carbon uptake.
- 417

418 Cement carbonation has historically not been included in national GHG inventories, since it has not been covered by the IPCC

- inventory guidelines and refinements. It is therefore technically, but not formally, excluded from the Paris Agreement. However,
 this is not from lack of interest from governments. Sweden has reported a Tier 1 cement carbonation calculation and memo in its
- 420 this is not non-fact of interest non-governments. Sweden has reported a free r cement carbonation carculation and meno-fi 421 National Inventory Report since 2020, but excludes this from its submitted inventory account. The UK has also recently
- 421 National inventory Report since 2020, but excludes this from its submitted inventory account. The OK has also recently
 422 published a Tier 2 methodology (DESNZ, 2023). The IPCC is expected to soon begin work on including a new chapter
- 422 published a Tiel 2 methodology (DESIV2, 2023). The free receives expected to soon begin work on including a new enapter 423 specifically on carbonation of both cement and lime in the Guidelines, as part of a new supplement to the 2006 Guidelines
- 423 spectrically on carbonation of both cement and line in the Orderlines, as part of a new supplement to the 2000 Orderlines
 424 (IPCC, 2025b). Once in the IPCC Guidelines, there is a clearer path for inclusion in national inventories. In third-party datasets
- 424 of rece, 2025b). Once in the rece of defines, there is a clearer pair for inclusion in national inventories. In unit-party datasets 425 of national emissions, cement carbonation is not included, but estimates of national totals have recently been published (Niu et
- 426 al., 2024).
- 427
- 428 One question that inclusion of cement and lime carbonation raises is whether it is a sink that can be directly and intentionally
- 429 modified, given that it is something that occurs to substances that have been already produced in the past (much like the concept





- 430 of "indirect anthropogenic effects" in the land use sector). Currently most concrete structures are designed to minimise
- 431 carbonation, because it often leads to structural weaknesses. Changes in design and cement composition (e.g. to allow the use of
- 432 aluminium reinforcing instead of steel) could lead to less focus on mitigating carbonation, and hence greater absorption of CO₂.
- 433 Further, at the end of a structure's life, when concrete is demolished, whether that concrete rubble is buried without access to air
- 434 or not has a large effect on further uptake, and this is something that could be controlled.

435 3 Conventions to assess emissions in different communities

- 436 For any given assessment of emissions, there are clearly many different decisions to be made regarding scope, system boundaries
- 437 and the selection of appropriate data. In this section, we therefore discuss how and why these choices are made in several
- 438 different assessment communities that are involved in estimating, tracking or using GHG emissions data. These different
- 439 conventions and their implied data choices are summarised in Fig. 7.

440 3.1 National targets, pledges and inventories under the UNFCCC

- 441 Countries are obligated to formulate climate targets, strategies and policies under the UNFCCC, most notably the Nationally
- 442 Determined Contributions (NDCs), which define their pledges to reduce emissions in the coming decades. The NDCs are one of
- the core mechanisms of the Paris Agreement and are formally linked to the national GHG inventories: where countries pledge
- 444 emissions reductions with reference to their historic emissions (e.g. a baseline level), that information should be in accordance
- with IPCC conventions and their national GHG inventories (UNFCCC, 2018a, 2022a). Similarly, national net zero targets are all
- based on inventory conventions. This ensures some degree of consistency in the agreement and encourages all countries to
- pledge emissions reductions in the same set of sectors, sources and gases, under the same definitions (e.g. of anthropogenic vs.
 natural sources and sinks).
- 449

450 A consequence of this for the research community and the IPCC is that independent assessments of current policy projections,

- the NDCs and net zero targets must also follow inventory conventions. This means that the data should cover the same scope of
- 452 sectors and gases as inventories, and that LULUCF estimates follow the inventory approach. Since ODS F-gases and cement
- 453 carbonation currently lie outside of the agreement, they must be excluded or treated separately. Similarly, it implies that fires not
- 454 occurring on managed land do not matter for target achievement, and that countries are able to define their approach to excluding
- atural disturbances (Australia and Canada have already done so in their NDCs with reference to their GHG inventory
- conventions, see above). Finally, international aviation and shipping emissions are something of a grey area under the agreement,
- 457 as they are reported as memos in national inventories, but are not accounted in totals nor towards national target achievement.
- Together, these requirements lead to a handful of options for tracking emissions: the PRIMAP Hist-CR dataset; the EDGAR
- dataset in combination with Grassi NGHGI (for CO₂ LULUCF), or national GHG inventories when dealing with individual or
 Annex I countries.
- 461

462 In practice, assessments of national pledges are mainly complicated by the LULUCF sector. The official synthesis of NDC

- 463 emissions projections prepared by the UNFCCC secretariat excludes inventory-based LULUCF emissions (UNFCCC, 2022b). In
- the scientific literature, emissions projections of the NDCs and current policies also tend to avoid the LULUCF sector
- 465 (Meinshausen et al., 2022), or carefully deal with it separately (Den Elzen et al., 2022). This is due to both ambiguity in the
- LULUCF contribution towards the targets of many countries (Fyson and Jeffery, 2019), as well as definitional differences
- 467 between the inventory and bookkeeping approach (section 2.2), which is consequential for benchmarking the NDCs against
- 468 integrated assessment models. By contrast, UNFCCC reporting on historical emissions includes LULUCF, following the national
- 469 inventory approach (UNFCCC, 2022c). The UNEP Emissions Gap Reports have also been reporting emissions at the national
- 470 level including inventory-based LULUCF for several years alongside a global total aligned with integrated assessment
- 471 modelling benchmarks (see next section) (UNEP, 2022, 2023, 2024). Indeed, detailed analyses of LULUCF pledges shows it is
- an important sector from the perspective of countries, both in their short and long-term targets (Grassi et al., 2017; Roman-
- 473 Cuesta et al., 2024). Thus while analyses focusing on national targets and pledges generally agree on the scope of emissions to
- 474 assess, approaches to LULUCF can differ.





475 **3.2 Integrated assessment modelling benchmarks**

Another key area where emissions assessments take place is in the integrated assessment modelling (IAM) literature. IAMs are used to derive future emissions scenarios under different assumptions of technology development and policy action, which then inform projections of climate change in the coming decades. IAMs therefore model future rather than past GHG emissions, but do so in accordance with a specific scope and set of system boundaries, and are often calibrated by or are compared to historic estimates of emissions (e.g. in terms of projected emission reduction rates). Notably, benchmarks from IAMs (e.g. describing emissions levels or reductions in the future that lead to different climate outcomes) are reported in the IPCC and are widely used to contextualise national or global progress towards the temperature goals of the Paris Agreement (IPCC, 2022; UNEP, 2024).

483

484 There are over a dozen different IAMs that regularly contribute to the literature, with significant heterogeneity in model structure 485 and scope. Nonetheless, they tend to follow several key conventions. First, most IAMs model the same basket of greenhouse 486 gases as national GHG inventories: the Kyoto gases. Emissions reporting is also often split into a similar set of sectors as in 487 national inventories and third-party emissions datasets (Byers et al., 2022). However, not all IAMs contain a land use model and 488 therefore some exclude the LULUCF sector. Beyond the Kyoto gases, many IAMs also include aerosols and other precursor 489 species with climate effects. IMAGE, MESSAGE-GLOBIOM, REMIND-MAgPIE and GCAM do not report ODS F-gases (The 490 common Integrated Assessment Model (IAM) documentation, 2025). Most IAMs do not incorporate cement carbonation, though 491 MESSAGE and IMAGE do. Second, and related to the LULUCF sector, IAMs typically only model anthropogenically induced 492 emissions. This follows from the basic purpose of IAMs, which is to analyse how human-driven technology and policy options 493 can shape the future climate response. Where a land use model is included, they predominantly consider direct anthropogenic 494 effects only associated with agriculture, land-use, land-use change or forestry (as in the Global Carbon Budget bookkeeping 495 approach), and do not include climate and fertilisation effects, particularly in forests remaining forests. Nonetheless, IAMs 496 continue to be improved to represent other emissions sources on land, both direct and indirect, including peatland rewetting, fire 497 emissions, and others.

498

499 Different IAM groups use different historical emissions data to calibrate their models, and the calibration is often done for 500 consistency and not for replication. There have been attempts to harmonise energy and emissions input data across models 501 (Giarola et al., 2021) and protocols written for various projects (Korsbakken et al., 2024), but because of differences in model 502 structure, it is not always possible or desirable to harmonise. To perform consistent climate assessments across multiple models it 503 is therefore necessary to post-process IAM emissions data, in-filling missing gases or sectors where necessary (Kikstra et al., 504 2022). Post-processed IAM results form the backbone of IPCC benchmarks of global action, such as the timing of net zero 505 emissions required to meet different climate objectives. To date these assessments have been aligned with the set of gases and 506 sectors outlined in inventories and the Paris agreement, but with the important difference that they use bookkeeping conventions 507 for the LULUCF sector. Other (e.g. non-Paris) emissions or land-based fluxes, as well as aerosols and precursor species, may be 508 modelled within IAMs, but these would be excluded from the benchmarks. This means that IAM benchmarks cannot be directly 509 compared to NDC assessments, but post processing can be used to translate between the two (Gidden et al., 2023; Grassi et al., 510 2021), as IAMs do not yet do this modelling natively. The translation and comparison of native IAM results to national inventory

- 511 definitions is already foreseen in the IPCC AR7 report outline. This is needed to emphasize the consequences of different
- 512 definitions and the fact that, after IAM results are translated to national inventory definitions, the carbon budget is reduced
- 513 (Gidden et al., 2023; Grassi et al., 2021), and reaching net-zero CO_2 alone will not suffice to prevent global warming (Allen et 514 al., 2025).
- . ,

515 **3.4 Climate forcing assessments**

516 A third area of emissions assessments takes the atmosphere as the primary frame of reference, in contrast to national accounts

517 and IAM benchmarks. The main objective of these is to track and explain observed changes to atmospheric concentrations of

518 greenhouse gases, to evaluate their influences on the climate (for instance through effective radiative forcing (ERF) estimates), or

- 519 to develop forward looking climate projections (Forster et al., 2024; Smith et al., 2021, 2024). They may also include emissions
- 520 budgeting studies and related publications that aim to track both natural and anthropogenic fluxes of emissions to the atmosphere
- 521 (Friedlingstein et al., 2025; Saunois et al., 2024; Tian et al., 2024; UNEP and CCAC, 2021). All are closely related to the climate
- 522 modelling literature.
- 523





524 Since these assessments aim to get the best estimate of GHG fluxes to the atmosphere, they would consider non-Paris Agreement 525 sources (ODS F-gases and cement carbonation) as well as all (non-CO₂) fire emissions. (Note that aerosol precursor species and 526 other short-lived climate forcers would also be considered relevant for these assessments). Removals due to natural sinks are 527 generally not considered as input data, as they are modelled directly by the climate models themselves, partly because these sinks 528 are functions of the climate state and hence are considered part of climate feedbacks. Additionally, climate modelling requires 529 inputs starting from pre-industrial, usually 17- or 1850, meaning that long time series datasets are often prioritised. Unless 530 studies are explicitly considering national boundaries or contributions to climate change (e.g. as in Jones et al., 2023), detailed 531 national or sectoral data is usually not required. This relaxes some constraints on using top-down observational datasets (e.g. fire 532 emissions observations, inversions of atmospheric concentrations), which often cannot be easily assigned to territorial 533 boundaries. These considerations lead to a few key sources being used for historical emissions: the GCB for CO₂, CEDS or 534 PRIMAP Hist-TP datasets for CO₂, CH₄ and N₂O emissions, the GFED dataset for CH₄ and N₂O from fires, and inversion 535 datasets (e.g. Velders et al., 2015; WMO, 2022) for F-gases (both UNFCCC and ODS).

536



537 538

Figure 7: Differences in greenhouse gas emissions conventions for three assessment communities. Box sizes are not representative of total
 emissions in each component.

541 4 Comparison and spread of GHG estimates

542 To what extent do the discussed issues of dataset coverage, definitions of anthropogenic emissions, and scope of the Paris

- 543 Agreement influence total GHG estimates?
- 544

545 In the first instance, only a handful of datasets come close to a complete coverage of inventory sectors and gases. For all Kyoto

546 gases (CO₂, CH₄, N₂O and UNFCCC F-gases) and excluding the LULUCF sector, these include PRIMAP Hist-TP, PRIMAP

547 Hist-CR and EDGAR. A fourth dataset covers these gases but excludes F-gases: CEDS. Between these datasets we observe

- 548 relatively minor deviations in total average decadal GHG emissions, the largest of which is due to differences in CH₄ estimates
- 549 between PRIMAP Hist-CR and PRIMAP Hist-TP (Fig. 7). For each of these datasets, the aggregate uncertainty range at a 90%
- 550 confidence interval is larger than the spread of values in other datasets.
- 551







Figure 8: Total Kyoto gas emissions across datasets, excluding LULUCF. Kyoto gases refer to CO₂, CH₄, N₂O and UNFCCC F-gas
emissions. Error bars indicate composite uncertainties of ±8 % for CO₂ (excl. LULUCF), ±30 % for CH₄ and F-gases, and ±60 % for N₂O,
corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6 WGI Chap.
7 (Forster et al., 2021). Note that the PRIMAP datasets exclude international bunker emissions.

557

558 Considering CO₂ emissions separately, we observe a low relative spread between datasets that cover a similar set of system

boundaries, but absolute differences of up to $1.7 \text{ GtCO}_2 \text{ yr}^{-1}$ (e.g. between the lowest estimate from PRIMAP Hist-CR and the

560 highest from EDGAR; Fig. 8). Relative differences as well as uncertainties are higher for CH₄ and N₂O emissions, with PRIMAP

561 Hist-CR - the PRIMAP time series that prioritises national inventory data - in particular reporting lower fossil CH₄ emissions

562 (Fig. 10 to 11). Indeed, several studies have pointed to relatively low estimates of fossil CH₄ in national inventories compared to

563 observational evidence (Deng et al., 2022; Janardanan et al., 2024; Scarpelli et al., 2022; Tibrewal et al., 2024).





564

Figure 9: Total carbon dioxide emissions across datasets, excluding LULUCF. Error bars indicate uncertainties of ±8 % for CO₂ (excl.
 LULUCF), corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6
 WGI Chap. 7 (Forster et al., 2021). Note that the PRIMAP datasets exclude international bunker emissions, and that the GCB estimate excludes
 cement carbonation.

569









570

Figure 10: Total methane emissions across datasets, excluding LULUCF. Error bars indicate uncertainties of ±30 % for CH4 (excl.
 LULUCF), corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR

LULUCF), corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6
 WGI Chap. 7 (Forster et al., 2021).

574

Total nitrous oxide emissions, excluding LULUCF Average annual Gt CO₂e, 2013-2022



575

579

By far the largest differences between datasets can be observed in the LULUCF sector (Fig. 12). According to decadal averages,
these range from net negative emissions of -3.44 to -2.85 GtCO₂e yr⁻¹ in the two national inventory aligned datasets (Grassi
NGHGI and PRIMAP Hist-CR), to net positive emissions in FAO (1.25 GtCO₂e yr⁻¹) and its derivative, PRIMAP Hist-TP (1.26

583 GtCO₂e yr⁻¹), to significantly larger net emissions in the GCB ($4.29 \text{ GtCO}_{2} \text{ yr}^{-1}$). As described in section 2.2, Grassi NGHGI

and GCB differ conceptually in terms of how they define anthropogenic removals and how they treat natural disturbances. And

as noted by Grassi et al. (2022a) the forest sink may be underestimated in FAOSTAT, because the underlying input data from

586 many developing countries (country reports to FAO-FRA) is incomplete. Of the datasets above, only some of them (FAOSTAT

and its derivative PRIMAP) include non-CO₂ emissions from fires and other land uses.

588

Figure 11: Total nitrous oxide emissions across datasets, excluding LULUCF. Error bars indicate uncertainties of ±60 % for N₂O (excl.
 LULUCF), corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6
 WGI Chap. 7 (Forster et al., 2021).









Figure 12: Total LULUCF emissions across datasets. Error bars indicate uncertainties of ±70 % for CO₂ LULUCF, ±30 % for CH₄ and ±60
 % for N₂O, corresponding to a 90 % confidence interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6
 WGI Chap. 7 (Forster et al., 2021).

593

594 Comparing emissions across the three assessment conventions outlined in section 3 (summarised in Fig. 7), significant

595 differences in total global greenhouse gas emissions can be observed (Fig. 13). Inventory-aligned emissions, shown here using

596 PRIMAP Hist-CR for non-LULUCF sectors and Grassi NGHGI for LULUCF, were 44.7 GtCO₂e yr⁻¹ [90% CI \pm 4.6] in the

597 decade 2014-2023. These emissions are both low relative to third-party datasets in terms of fossil methane, exclude international

aviation and shipping, and include the inventory-aligned definition of LULUCF. This is ~ 8.4 GtCO₂e yr⁻¹ lower than emissions

599 comparable with IAM benchmarks, primarily due to the bookkeeping definition of LULUCF (\sim 7.1 GtCO₂e yr⁻¹), but also due to

600 lower estimates of fossil methane in inventory prioritised data, as well as the inclusion of bunker emissions (the latter adding 1.1

601 GtCO₂ yr⁻¹ between 2014-2023). Expanding the scope further to consider non-Paris sources (ODS F-gases, cement carbonation)

and all global fire emissions of CH_4 and N_2O , decadal average emissions increase by 1.8 GtCO₂e yr⁻¹, of which 1.87 GtCO₂e yr⁻¹

is from ODS F-gases, -0.72 GtCO₂ yr⁻¹ is from the cement carbonation sink, and 0.64 GtCO₂ yr⁻¹ is from fires. Interannual variability in emissions also increases due to large fluctuations in annual fire emissions (Fig. 4).

605

The emissions shown in Figure 13 do not include non-CO₂ emissions in the LULUCF sector, which based on FAOSTAT would be $0.25 \text{ GtCO}_2 \text{ yr}^{-1}$. Further, none of the datasets in Figure 13 (nor FAOSTAT) include the indirect anthropogenic portion of

emissions from wetlands or freshwater bodies, which aggregated and estimated from individual studies may sum to as much as

607 608 609

2.4 GtCO₂ yr⁻¹ as discussed in section 2.2.3.

610



Differences in total greenhouse gas emissions Average annual Gt CO₂e, 2014-2023





- 612 **Figure 13: Differences in total greenhouse gas emissions under different data and system boundary choices.** Error bars indicate
- 613 uncertainties of ±8% for CO₂ Fossil, ±70 % for CO₂ LULUCF, ±30 % for CH₄ and ±60 % for N₂O, corresponding to a 90 % confidence
- 614 interval following Minx et al. (2021). CO₂e emissions are calculated using GWP100 from AR6 WGI Chap. 7 (Forster et al., 2021). Data: for
- 615 'inventory-aligned emissions' PRIMAP Hist-CR (non-LULUCF sectors) and Grassi NGHGI (LULUCF) (Grassi et al., 2022a; Gütschow et al.,
- 616 2025); for 'IAM benchmark-aligned emissions' GCB (all CO2 incl. LULUCF) and PRIMAP Hist-TP (non-CO2, excl. LULUCF)
- 617 (Friedlingstein et al., 2025; Gütschow et al., 2025); for 'Additional sources' Forster et al. (ODS F-gases), GFED (fires) and GCB (cement
- carbonation) (Forster et al., 2024; Friedlingstein et al., 2025; van der Werf et al., 2017).

619 5 Discussion and conclusion

- 620 In this article we have explored key reasons why GHG emissions estimates differ, namely that datasets vary in their coverage of
- 621 gases, sectors and countries; that there are different approaches to defining 'anthropogenic' emissions; and that the Paris
- 622 Agreement doesn't cover all relevant sources of emissions. Importantly, we find that there are multiple possible approaches to
- addressing these issues, and that these depend on different decision criteria determining the scope and conceptual boundaries of
- an assessment. Among the assessment conventions we have described, such criteria include *political relevance* (where an
- assessment aims to be consistent with the scope of the Paris Agreement), that emissions should be *direct anthropogenic only*
- 626 (where an assessment excludes sources and sinks that are less amenable to direct policy intervention), or that emissions should be
- 627 accurate compared to observations (where an assessment aims to describe the best estimate of fluxes consistent with
- 628 observations). Other decision criteria are also possible, some of which are mutually exclusive or in conflict with one another
- 629 (Table 2). This underlines the importance of clearly stating which criteria drive an assessment, and what they imply in terms of 630 emissions coverage and system boundaries.
- 631
- For some components of emissions, it is straightforward to quantify the impact of including or excluding them from totals. This is the case for ODS F-gases, cement carbonation, as well as for the LULUCF sector where significant efforts have been made to
- explain differences and provide translation methodologies between estimates (Friedlingstein et al., 2025; Grassi et al., 2023;
- 635 Schwingshackl et al., 2022). However, for others the impact of different conceptual approaches is more challenging to quantify.
- For instance, while the broad treatment of fire emissions in inventories, models and third-party datasets is known (Fig. 3),
- 637 quantifying these differences would require directly comparing their estimates of burned areas and emissions within the
- 638 LULUCF sector. While this is largely available in the national GHG inventories, these are globally incomplete. Further,
- 639 observational datasets such as GFED do not differentiate by national borders; while others (e.g. FAO, GCB, PRIMAP-Hist) do.
- 540 Similarly, in the case of wetlands and freshwater bodies, there are estimates in literature on global fluxes, but little work on
- 641 comparing these to bottom-up, national or inventory estimates although such comparisons have been made for N₂O (Conchedda
- and Tubiello, 2020). As a result, differences in how datasets treat indirect anthropogenic fluxes from fires and wetlands are
- 643 largely unknown to non-domain experts. As interest grows in the potentials, limits and risks of carbon dioxide removal and
- 644 "natural climate solutions" including wetland restoration (Ma et al., 2024; Zou et al., 2022), it may become increasingly
- 645 important to assess these fluxes with more specificity.
- 646
- 647 Overall, we find significant differences between global GHG estimates, primarily driven by the LULUCF sector, but also with 648 non-trivial impacts from including non-Paris Agreement sources. Nonetheless, emissions are unambiguously increasing and are 649 far off track from levels and trends consistent with meeting the objectives of the Paris Agreement. At a national level, even larger 650 relative differences are to be expected for countries with significant land or forest areas. As it stands, though, we lack sufficient
- and comprehensive national data for ODS F-gases and fires to evaluate their influence below the global scale, though individual
- 652 studies are starting to fill this gap (Niu et al., 2024).
- 653

Relevant decision criteria	Description	Example use cases
Pragmatism	Datasets are chosen that are relatively up-to-date and complete, open source and easily machine readable, provide a reasonable level of national or sectoral detail, but do not necessarily use high tier estimation methods	Ex-post evaluations of climate policy effectiveness (Stechemesser et al., 2024)



Political relevance	Datasets should be officially recognised by parties to the UNFCCC, can be estimated by countries with varying institutional capacities, and are consistent with those used to inform national emission reduction pledges	Evaluations of implied emissions reductions under climate pledges, the NDCs or national net zero targets (Den Elzen et al., 2022)
Consistency	Datasets are chosen to be consistent with the frameworks and uses of the respective assessment community, for example to harmonise with definitions of "net zero", or use consistent assumptions as taken in the IPCC	Updates of IPCC indicators (Forster et al., 2024)
Direct anthropogenic only	Datasets and their sources are conceptually limited to only the set of activities that are directly human driven and thereby amenable to policy intervention	Integrated assessment modelling benchmarks and bookkeeping land use change models (UNEP, 2024)
Accuracy compared to observations	The group of sources, including indirect anthropogenic emissions, that gives the best estimate of the flux to the atmosphere compared to observations	Greenhouse gas budgeting studies (Friedlingstein et al., 2025; Saunois et al., 2024; Tian et al., 2024)
Time series since pre- industrial	Analysis is dependent on a time series since pre-industrial (e.g. 1750 or 1850)	Modelling of historic contributions to climate change (Jones et al., 2023)

654

Table 2: Decision criteria for selecting and using emissions data.

655

A multitude of activities and processes drive GHG emissions, many of which interact with natural systems. The resulting data is

therefore inherently complex, with nuances that may not be obvious to users lacking specific domain knowledge. Some issues

also cannot be resolved easily, such as the attribution of synergistic effects of anthropogenic and natural drivers. Despite this,

659 GHG emissions data is very widely used and remains one of the most important indicators of human impact on the planet.

Different choices of data can have wide reaching implications, especially at a national level where varying definitions (e.g. of

661 LULUCF emissions) could cast doubt over a country's claimed mitigation progress. We have therefore attempted to explain

some of the key factors that drive differences between estimates, as well as the decision criteria underlying these choices. We

recommend that data users familiarise themselves with these issues, and take steps to clearly state the decision criteria behind

their own choices and what impact it may have on their analysis.

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667 (Lamb, 2025b). 668 669 Code availability: the code used in this study to make figures 8-13 is available at: https://github.com/ClimateIndicator/GHG-670 Emissions-Assessment/blob/main/GHG-Emissions-Assessment-Differences-2025.Rmd 671 672 Author contribution: WFL, RMA, GPP, CS, JP, PF, JM and PMF conceptualised the study. WFL conducted the analysis and 673 prepared figures. All authors contributed to writing, reviewing and editing the draft. 674 675 Competing interests: Author FNT is a member of the editorial board of the journal. 676 677 Funding acknowledgements: RMA and GPP were supported by the Horizon Europe Research and Innovation Programs under 678 grant agreement nos. 101081395 (EYE-CLIMA) and 101056907 (Pathfinder). WFL was supported by the European Union ERC-679 2020-SyG 'GENIE' (951542) grant. CS was supported by the Horizon Europe Research and Innovation Programs under grant 680 agreement nos. 101081661 (WorldTrans) and 101081369 (SPARCCLE). ZN acknowledges funding from the European Union's 681 Horizon 2020 research and innovation programmes (grant agreement no. 101003536) (ESM2025). 682

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